Bioaccumulation of ¹⁰⁶Ru by marine Phytoplankton

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ABSTRACT

The chlorophyte <u>Dunaliella tertiolecta</u> and the diatom <u>Thalassiosira pseudonana</u> readily accumulate $106\frac{1}{Ru}$ reaching volume concentration factors of roughly 10^3 and 10^4 , respectively. The uptake process is passive and is most likely related to surface sorption. High uptake by live and dead cells (phytodetritus) indicate that phytoplankton were likely initial vectors in the rapid vertical transport of $106\frac{100}{Ru}$ noted after the arrival of Chernobyl fallout.

INTRODUCTION The radioisotopes of ruthenium, particularly ¹⁰³Ru and ¹⁰⁶Ru, are considered among the more important radioactive contaminants because of their relatively high yield from fission and their moderately long half-lives. Ruthenium-106 (T $_{\rm X}^2 \propto 373$ days) has entered the marine environment primarily as fallout from previous nuclear tests and in waste effluents from several nuclear reprocessing plants. Furthermore, the recent accident at Chernobyl resulted in a major input of ¹⁰⁶Ru to marine waters (Powler et al., 1987). The behaviour of ruthenium in sea water is complex (IAEA, 1975) and a large fraction of the radionuclide concentration is associated with particulate matter (Coughtrey and Thorne, 1983). The relatively high reactivity of ¹⁰⁶Ru results in its being readily accumulated by a variety of marine organisms including phytoplank-ton. Data from recent Chernohyl fallout studies (Fowler et al., 1987; Kempe and Nies, 1987) have suggested that phytoplankton is probably the vector by which ¹⁰⁶Ru enters the pelagic food chain and is subsequently transported vertically through the water colum. Since information on the mechanisms controlling ¹⁰⁶Ru uptake by phyto-plankton species is limited, a series of experiments were undertaken to examine ¹⁰⁶Ru biokinetics in two common species of phytoplankton.

METHODS AND MATERIALS

METHODS AND MATERIALS The chlorophyte <u>Dunaliella tertiolecta</u> (clone DUN) and the diatom <u>Thalassiosira</u> <u>pseudonana</u> (clone 3N) were used in all experiments. Different concentrations of cells in mid to late log phase were exposed to ¹⁰⁶Ru (as RuCl₂) in the light and dark under rigorously controlled conditions similar to those described previously for other radionuclides (Fisher <u>et al.</u>, 1983). In addition, heat-killed cells of both species were exposed in parallel with live cultures to examine the effect of metabo-lism on uptake. All culture conditions filtration techniques, counting procedures and computation of volume concentation factors were identical to those employed by Fisher <u>et al</u>. (1983) for these species.

RESULTS AND DISCUSSION

RESULTS AND DISCUSSION The results of two sets of experiments each carried out in triplicate are pre-sented in Table 1. It is clear that uptake by these two species is primarily a pas-sive adsorptive process since there was virtually no difference in the accumulation pattern between live cells in the dark and dead cells. Furthermore, during the first 24 hours, uptake by live cells in the light was also identical to the other two treatments; however, between days 1 and 3 concentration factors decreased as cell density increased. This is most likely an effect of reduced uptake by biological dilution. Uptake was also noticeably higher in the green alga <u>Dunaliella</u> (VCF $\approx 10^5$)

Table 1. Bioaccumulation of 106 Ru (Volume concentration factor x 10^4)⁺ over time in uptake experiments with <u>D. tertiolecta</u> (DUN) and <u>T. pseudonana</u> (38).

Species	Treatment	1 hr.	1 d.	2 d.	3 d.				
(VCF x 10 ⁴)									
3H	A L	0.174	2.55	1.72	1.10				
3H	A D	0.112	5.65	9.85	14.0				
3H	De	0.23	4.17	15.3	37.6				
DUN	AL	0.95	36.8	25.1	16.8				
DUN	A D	0,72	46.7	81.8	122				
DUN	De	1.34	47.1	92.7	118				

VCF values are averages of two experiments with three replicates each.

than in the diatom <u>Thalassiosira</u> (VCF $\leq 10^4$). This differs from the transuranic elements plutonium and americium with which the reverse is found (Fisher <u>et al.</u>, 1983). In any event, the relatively high VCPs for 106Ru in these two species, which approach those reported for transuranics, indicate that 106Ru entering the sea would rapidly become associated with phytoplankton which could then be passed on to xoo-plankton grazing the cells. Thus, contaminated phytodetritus and xooplankton excreta would become prime vectors for rapidly moving ruthenium downward in the water column as was seen to occur immediately following the Chernobyl accident in both the Mediterranean (Fowler <u>et al.</u>, 1987) and the North Sea (Kempe and Nies, 1987).

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Transfer of 137Cs across the Gills epithelial cells of the Crab Carcinus mediterraneus

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Apart from their function in osmoregulation and respiration the gills of marine organisms play an absorptive role in distributing a wide range of pollutants into the various tissues and organs. The role of the gills in radioactive caesium transport and distribution in Grustacea has been pointed out by Bryan (1961a, 0), but the problem has not yet been solved in detail. Therefore, for a better understanding the transport of radioactive pollutant in epithelial gill cells we studied the transfer of Cs-137 across perfused isolated <u>Carcinus</u> gill preparations. Green crabs <u>Carcinus</u> mediterraneus Czrn., were obtained from the central Adriatic pear the seland of Dugl Otok. Crabs were acclimated in diluted sea water (15 x 10 -salinity), fed and kept at a constant room temperature (16 °C). In the experiments described 3 posterior gill pairs were collected for the perfusion technique. Experimental details were described by Lucu and Siehers, 1986; 1987. Fadioactive caesium (3.7 KBg Cs-137/ml diluted sea water) was used for experimental dilary tubes and preparation was fixed by neoprene block immersed in the bathing solution (20 - 40 ml). Identical diluted sea water (260 ml Na) to the bathing solution (20 - 40 ml). Identical diluted sea water (260 ml Na) to the bathing solution (20 - 40 ml).

Transport factor= $\frac{137_{CS} \text{ transported across epithelia}}{137_{CS} \text{ in bathing solution x W}}$ (1)

where W is the fresh gill weight in grams.

where W is the fresh gill weight in grams. Effects of diuretic amiloride (Merch Sharp Dohme, Munich) on caesium fluxes from the apical (bathing side) to the basolateral haemolymph side (influxes) and in the opposite direction (effluxes) were studied (Table 1.), Cs-137 and stable Cs effluxes are greater than influxes, showing that the basolaterally oriented gill side is more permeable to Cs-137 than the apically oriented one. Therefore, the Cs-137 transport factors are larger from the haemolymph side to the bathing side (effluxes) than the fluxes in the opposite direction (Table 1.). In addition, influxes and effluxes of Cs-137 (and stable caesium) was inhibited by amiloride (O.1 mM), added to the bathing solution. This relatively high concentration of amiloride showed a similar effect on Na' inhibition from the apical site as reported by Lucu and Siebers, 1966, Since Cs have similar physioc-chemical behaviour to K (Bryan, 1961) we suggest that Cs competes with K' for the K'/H' exchanger located on the apical membrane side. Since amiloride failed to affect this mechanism at the lower concentrations the presence of an ion channel interaction was precluded. Transport factor TBP Caesium fluxe

Transport factor			TBP	Caesium lluxe	
	(% of ¹³⁷ Cs)	r 0.1 g ⁻¹ h ⁻¹)	(mV)	(jumol x g	$^{-1}$ h ⁻¹)
Control	1.65±0.20	6.49 ± 1.04	-3.1	0.99 <u>+</u> 0.09	2 . 28 ± 0.70
Amiloride	0.93 [±] 0.11	4.70±0.46	-7.5	0.56±0.01	1.51±0.01

Table 1. Caesium fluxes through the isolated perfused Carcinus gill epithelia. Mean values of 6 observations $\frac{1}{4}$ S.E. TBPs transbranchial potential (in mV).

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